



## New insight into the atmospheric chloromethane budget gained using gained using

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chloromethane  
budget

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# New insight into the atmospheric chloromethane budget gained using stable carbon isotope ratios

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## Abstract

Atmospheric chloromethane ( $\text{CH}_3\text{Cl}$ ) plays an important role in stratospheric ozone destruction, but many uncertainties still exist regarding strengths of both sources and sinks and the processes leading to formation of this naturally occurring gas. Recent work has identified a novel chemical origin for  $\text{CH}_3\text{Cl}$ , which can explain its production in a variety of terrestrial environments: The widespread structural component of plants, pectin, reacts readily with chloride ion to form  $\text{CH}_3\text{Cl}$  at both ambient and elevated temperatures (Hamilton et al., 2003). It has been proposed that this abiotic chloride methylation process in terrestrial environments could be responsible for formation of a large proportion of atmospheric  $\text{CH}_3\text{Cl}$ . However, more information is required to determine the global importance of this new source and its contribution to the atmospheric  $\text{CH}_3\text{Cl}$  budget.

A potentially powerful tool in studying the atmospheric  $\text{CH}_3\text{Cl}$  budget is the use of stable carbon isotope ratios. In an accompanying paper it is reported that the reaction of  $\text{CH}_3\text{Cl}$  with OH radical, the dominant sink for atmospheric  $\text{CH}_3\text{Cl}$ , is accompanied by an unexpectedly large fractionation factor (Gola et al., 2005). Another recently published study shows that  $\text{CH}_3\text{Cl}$  formed by the abiotic methylation process at ambient temperatures has a unique stable carbon isotope signature, extremely depleted in  $^{13}\text{C}$ , unequivocally distinguishing it from all other known sources (Keppler et al., 2004). Using these findings together with data existing in the literature, we here present three scenarios for an isotopic mass balance for atmospheric  $\text{CH}_3\text{Cl}$ . Our calculations provide strong support for the proposal that the bulk fraction of atmospheric  $\text{CH}_3\text{Cl}$  ( $1.8$  to  $2.5 \text{ Tg yr}^{-1}$ ) is produced by an abiotic chloride methylation process in terrestrial ecosystems, primarily located in tropical and subtropical areas, where turnover of biomass is highest. Furthermore our calculations also indicate that the microbial soil sink for  $\text{CH}_3\text{Cl}$  is likely to be much larger ( $>1 \text{ Tg yr}^{-1}$ ) than that previously assumed.

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1. Introduction

Chloromethane (CH<sub>3</sub>Cl) is the most abundant halocarbon in the atmosphere with a reported average mixing ratio in the range of 530 to 560 pptv (Montzka and Fraser, 2003; Simmonds et al., 2004; Trudinger et al., 2004; Aydin et al., 2004; Yoshida et al., 2004), corresponding to a total atmospheric burden of around 4 to 5 Tg (tera gram = 10<sup>12</sup> gram). Although largely of natural origin CH<sub>3</sub>Cl is responsible for around 16% of chlorine-catalysed ozone destruction in the stratosphere (Montzka and Fraser, 2003). With large reductions in anthropogenic emissions of many ozone-depleting gases mandated under the Montreal Protocol, halogenated gases with natural sources will become relatively more important as a source of chlorine in the stratosphere in the future. A better understanding of the atmospheric budget of chloromethane is therefore required for reliable prediction of future ozone depletion.

Until 1996 most of the CH<sub>3</sub>Cl input to the atmosphere was considered to originate from the oceans, but investigations in recent years have clearly demonstrated that terrestrial sources dominate the atmospheric budget (Moore et al., 1996; Harper and Hamilton, 2003; Montzka and Fraser, 2003). The latest consensus assessment by the World Meteorological Organization (WMO) (Montzka and Fraser, 2003) considered the major terrestrial sources of CH<sub>3</sub>Cl to be biomass burning (Lobert et al., 1999; Andreae and Merlet, 2001), wood-rotting fungi (Watling and Harper, 1998), coastal salt marshes (Rhew et al., 2000), and tropical vegetation (Yokouchi et al., 2002). In addition decomposition of soil organic matter can act as a source for atmospheric CH<sub>3</sub>Cl (Keppler et al., 2000), but the importance of this source remains uncertain. Despite the progress in identifying new sources, the best estimate in the latest WMO Report (Montzka and Fraser, 2003) still revealed a significant shortfall of >1 Tg yr<sup>-1</sup> between known sources and the modelled sinks. However, very recently Hamilton et al. (2003) made the interesting discovery that the widespread plant structural component, pectin, reacts readily with chloride ion to form chloromethane. Furthermore it was established that the reaction occurs abiotically at ambient temperatures in senescent and weath-

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ering plant material. They calculated that this process acting on senescent leaves and leaf litter could be responsible for formation of a large proportion of atmospheric chloromethane (up to  $2.5 \text{ Tg yr}^{-1}$  from tropical and sub-tropical regions) and may also provide the mechanism for release of chloromethane during biomass burning, possibly for emissions from living tropical vegetation and importantly also the abiotic soil emissions reported by Keppler et al. (2000). Including this new emission source allows global  $\text{CH}_3\text{Cl}$  sources to balance and even possibly outweigh the best estimates of total sinks. Table 1 summarises existing information on estimated emission fluxes from known sources. The estimates of the strengths of many sources have large uncertainties. For example the possible  $\text{CH}_3\text{Cl}$  flux from tropical ferns and trees ranges from 0.8 to  $8.2 \text{ Tg yr}^{-1}$  (Yokouchi et al., 2002) whilst  $\text{CH}_3\text{Cl}$  emissions in the tropics and sub-tropics by senescent leaves and leaf litter ranging from 0.03 to  $2.5 \text{ Tg yr}^{-1}$  have been suggested (Hamilton et al., 2003).

As regards sinks for atmospheric  $\text{CH}_3\text{Cl}$ , the dominant loss process for atmospheric  $\text{CH}_3\text{Cl}$  is via reaction with photochemically produced OH radicals (Table 2) and is considered responsible for some  $3.2 \text{ Tg yr}^{-1}$  (Montzka and Fraser, 2003). The reaction of  $\text{CH}_3\text{Cl}$  with chlorine radicals in the marine boundary layer constitutes another significant sink that has been estimated at  $370 \text{ Gg yr}^{-1}$  (giga gram =  $10^9$  gram). Small proportions of tropospheric  $\text{CH}_3\text{Cl}$  are lost to the stratosphere ( $200 \text{ Gg yr}^{-1}$ ) and to mid- and high- latitude waters ( $75 \text{ Gg yr}^{-1}$ ). Another sink for atmospheric  $\text{CH}_3\text{Cl}$ , is degradation by soil microorganisms for which Montzka and Fraser (2003) gave a best estimate of  $0.18 \text{ Tg yr}^{-1}$ . However we believe that compelling evidence is now emerging from microbial studies for a much larger soil sink for  $\text{CH}_3\text{Cl}$ . The ubiquitous occurrence in soil from pristine environments of microbial species capable of utilizing  $\text{CH}_3\text{Cl}$  as sole carbon and energy source implies a widespread and substantial soil sink for  $\text{CH}_3\text{Cl}$  (Harper, 2000; McAnulla et al., 2001; Harper and Hamilton, 2003; Miller et al., 2004). If the microbial sink for  $\text{CH}_3\text{Cl}$  in soil is assumed to account for a similar proportion of the atmospheric burden as with  $\text{CH}_3\text{Br}$ , the magnitude of the soil sink would be of the order  $1.6 \text{ Tg yr}^{-1}$  (Harper and Hamilton, 2003). Since the atmospheric concentration

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of  $\text{CH}_3\text{Cl}$  is around 60 fold higher than that of  $\text{CH}_3\text{Br}$  it may well act as a far better substrate for microorganisms than atmospheric  $\text{CH}_3\text{Br}$ . Hence even a soil sink of  $1.6 \text{ Tg yr}^{-1}$  for  $\text{CH}_3\text{Cl}$  may be a considerable underestimate. For the purpose of this paper the values of the oceanic source and sink quoted by Montzka and Fraser (2003) will be used. These numbers do in fact represent a net source in warm waters and a net sink in colder waters (the temperature threshold between these waters being around  $12^\circ\text{C}$ ). A rigorous isotopic budget would take into account the fact that the gross sources and sinks are larger, the loss processes, for example, not being restricted to cold waters.

A potentially powerful tool in the investigation of the atmospheric  $\text{CH}_3\text{Cl}$  budget is the use of stable carbon isotope ratios ( $^{13}\text{C}/^{12}\text{C}$ ). For example, this technique has been applied with some success to investigate sources and sinks of atmospheric methane (Miller et al., 2002; Quay et al., 1999) carbon monoxide (Röckmann et al., 1998, 2002) and nitrous oxide (Kim and Craig, 1993; Röckmann et al., 2003) (see Gros et al., 2004 for further references). Furthermore  $^{13}\text{C}/^{12}\text{C}$  ratio measurements are useful in studying non-methane volatile organic compounds (VOCs) in atmospheric samples (Rudolph et al., 1997; Tsunogai et al., 1999). Measurements of the  $^{13}\text{C}/^{12}\text{C}$  isotopic ratio of VOCs including  $\text{CH}_3\text{Cl}$  are usually reported with respect to a standard in “delta” notation where  $\delta^{13}\text{C}$  is defined as:

$$\delta^{13}\text{C}(\text{‰}) = (R_{\text{sample}}/R_{\text{standard}} - 1) \times 1000 \text{ ‰} \quad (1)$$

with  $R = ^{13}\text{C}/^{12}\text{C}$ , where  $R_{\text{standard}}$  is the  $^{13}\text{C}/^{12}\text{C}$  ratio of an international standard, usually Vienna Pee Dee Belemnite (V-PDB), with a value of  $R_{\text{standard}} = 0.0112372$ .

In this paper using an isotopic mass balance approach we attempt to shed light on the question of which  $\text{CH}_3\text{Cl}$  source in the terrestrial environment provides the bulk fraction to the atmosphere. We have utilised recently published information on the  $\delta^{13}\text{C}$  values of  $\text{CH}_3\text{Cl}$  sources and the latest measurements for the kinetic isotope effects (KIEs) associated with sink processes to determine which source/sink distribution pattern provides the best fit with the observed carbon isotope ratio of atmospheric  $\text{CH}_3\text{Cl}$ .

## 2. Stable isotope composition of known sources of tropospheric CH<sub>3</sub>Cl

The isotope signatures of several CH<sub>3</sub>Cl sources have been measured. Values of  $\delta^{13}\text{C}$  for CH<sub>3</sub>Cl released during biomass burning were first measured by Rudolph et al. (1997) and subsequently by Czapiewski et al. (2001). Both studies showed that CH<sub>3</sub>Cl emitted is relatively depleted in  $^{13}\text{C}$  with  $\delta^{13}\text{C}$  values ranging from  $-38\%$  to  $-68\%$ . For instance, burning of wood of manuka (*Leptospermum scoparium*) resulted in a  $\delta^{13}\text{C}$  for CH<sub>3</sub>Cl released of  $-45.1 \pm 0.6\%$  (Rudolph et al., 1997) which is considerably depleted compared with the unburnt fuel ( $\Delta = -17\%$ ; in this paper we define the isotope difference between two pools as  $\Delta = \delta^{13}\text{C}_{\text{pool1}} - \delta^{13}\text{C}_{\text{pool2}}$ ). Combustion of wood of *Eucalyptus* spp. yielded  $\delta^{13}\text{C}$  values for CH<sub>3</sub>Cl emissions of  $-51.7\%$  (Czapiewski et al., 2001) which was also highly depleted in comparison with unburnt fuel ( $\Delta = -24.8\%$ ). Globally the bulk of biomass burning involves C<sub>3</sub> plant material but approximately one-third is associated with tropical and sub-tropical grasslands which are dominated by C<sub>4</sub> plants (Harper et al., 2001). Thompson et al. (2002) assumed that the isotopic fractionation of CH<sub>3</sub>Cl relative to the parent fuel is similar for the combustion of material from both C<sub>3</sub> and C<sub>4</sub> plants ( $\Delta = -25 \pm 12\%$ ) and suggested an average  $\delta^{13}\text{C}$  for CH<sub>3</sub>Cl emissions during biomass burning of  $-47 \pm 12\%$  although this has not been confirmed empirically.

It has been argued that CH<sub>3</sub>Cl release by tropical ferns and trees such as dipterocarps could represent the single largest global source of atmospheric CH<sub>3</sub>Cl (Yokouchi et al. 2002) although no mechanism has been suggested. Values for  $\delta^{13}\text{C}$  for CH<sub>3</sub>Cl released from two species of glasshouse-grown tropical ferns (*Cyathea smithii* and *Angiopteris evecta*) of  $-72.7 \pm 1.4\%$  and  $-69.3 \pm 0.9\%$  reported by Harper et al. (2003) showed considerable depletion ( $\Delta = 42.3$  and  $43.4\%$ ) relative to bulk biomass. This is consistent with the average value obtained for three families of tropical ferns reported by Komatsu et al. (2004). An even greater  $^{13}\text{C}$  fractionation has been observed in CH<sub>3</sub>Cl emissions by leaves of a glasshouse-grown dipterocarp (*Shorea guiso*) for which a  $\delta^{13}\text{C}$  value of  $-75.2 \pm 1.7\%$  was determined by Hamilton, Yokouchi, Yukawa

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and Harper (unpublished results). Harper et al. (2001) measured the  $\delta^{13}\text{C}$  of  $\text{CH}_3\text{Cl}$  released by different batches of tubers of the potato (*Solanum tuberosum*). The overall mean  $\delta^{13}\text{C}$  for  $\text{CH}_3\text{Cl}$  emissions was  $-62.2 \pm 4.6\text{‰}$ , again showing a substantial depletion ( $\Delta = 34.5\text{‰}$ ) relative to the bulk biomass of the tubers. The same group also reported  $\delta^{13}\text{C}$  values of  $\text{CH}_3\text{Cl}$  released by glasshouse-grown material of the halophyte *Batis maritima* of  $-65.7 \pm 3.4\text{‰}$ , with significant depletion ( $\Delta = 36.8\text{‰}$ ) observed relative to biomass. Later in situ studies on  $\text{CH}_3\text{Cl}$  emissions by *Batis maritima* in Californian salt marsh largely confirmed these observations with weighted daily mean  $\delta^{13}\text{C}$  values of  $-62 \pm 3\text{‰}$  recorded (Bill et al., 2002).

Wood rotting fungi are another significant terrestrial source of atmospheric  $\text{CH}_3\text{Cl}$  (Watling and Harper, 1998). Laboratory studies on the fungus *Phellinus pomaceus* cultured on wood showed  $^{13}\text{C}$  depletion in  $\text{CH}_3\text{Cl}$  released relative to the wood substrate. The  $\delta^{13}\text{C}$  measured for this source was  $-43.3 \pm 0.2\text{‰}$ , a fractionation of  $17.9\text{‰}$  relative to the substrate (Harper et al., 2001).

$\delta^{13}\text{C}$  values for industrially produced  $\text{CH}_3\text{Cl}$  from various sources have been variously reported as  $-41.9 \pm 0.3\text{‰}$  (Rudolph et al., 1997),  $-58.4 \pm 0.3\text{‰}$  (Holt et al., 1997) and for two different batches  $-46.9 \pm 0.8\text{‰}$  and  $-61.9 \pm 0.5\text{‰}$  (Harper et al., 2001). The isotopic composition of  $\text{CH}_3\text{Cl}$  emissions from coal combustion and incineration has not yet been measured. However as coals have a  $^{12}\text{C}/^{13}\text{C}$  distribution similar to land plants (Hoefs, 1980) we assume that the  $\text{CH}_3\text{Cl}$  fractionation from both sources is the same as that calculated for biomass burning. Any uncertainty regarding this value will not have a significant effect on the atmospheric  $\text{CH}_3\text{Cl}$   $^{13}\text{C}$  signature as both sources combined contribute only about 3% to the total  $\text{CH}_3\text{Cl}$  budget.

The most abundant  $\text{C}_1$  units of terrestrial plants, the methoxyl groups ( $\text{OCH}_3$ ) of pectin and lignin, which are responsible for  $\text{CH}_3\text{Cl}$  emissions from senescent leaves and leaf litter referred to earlier in this paper, have a unique carbon isotope signature exceptionally depleted in  $^{13}\text{C}$  (Keppler et al., 2004). The carbon isotope signatures of pectin and lignin methoxyl groups of leaf tissue from trees, grasses and halophytes including plants from  $\text{C}_3$ ,  $\text{C}_4$  and CAM plant categories were found to be in the range

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of  $-33$  to  $-77\%$  and the depletion of the plant methoxyl pool relative to bulk plant biomass ranged from  $-11$  to  $-46\%$  with the pectin  $C_1$  pool generally more depleted than the lignin  $C_1$  pool. Keppler et al. (2004) reported that  $CH_3Cl$  released during low temperature heating of plant leaf tissue was even more depleted in  $^{13}C$  than either of the  $C_1$  plant methoxyl pools. They found  $\delta^{13}C$  values of  $-147$ ,  $-139$  and  $-119\%$  for  $CH_3Cl$  emitted at  $40^\circ C$  from dried leaf tissue of two tree species, European ash (*Fraxinus excelsior*) and wych elm (*Ulmus glabra*), and the grass species, cocksfoot (*Dactylis glomerata*), respectively. These  $\delta^{13}C$  values are the lowest ever observed in a terrestrial carbon compound produced by natural processes and represent a striking depletion of  $119$ ,  $108$  and  $89.9\%$  relative to the respective biomass. A large kinetic isotope effect (KIE) is obviously involved in the process leading to  $CH_3Cl$  release during heating of such plant material. At the higher temperature of  $225^\circ C$  these researchers also reported very low  $\delta^{13}C$  values for  $CH_3Cl$  emissions from a variety of  $C_3$  plants ( $-89.7 \pm 10.3\%$ ,  $\Delta = 60.2\%$  relative to biomass), a  $C_4$  plant, maize (*Zea mays*,  $-91.3\%$ ,  $\Delta = 80.3\%$  relative to biomass) and for two CAM plants, saltwort (*Batis maritima*,  $-78.3\%$ ,  $\Delta = 52.7\%$  relative to biomass) and scarlet paintbrush (*Crasula falcata*,  $-81.4\%$ ,  $\Delta = 63.5\%$  relative to biomass). Clearly therefore, the depletion of  $^{13}C$  in  $CH_3Cl$  released during heating of leaf tissue is widespread amongst all plant categories, although this depletion was found to decrease as the heating temperature increased.

The only  $\delta^{13}C$  values for  $CH_3Cl$  from an oceanic source are those reported by Komatsu et al. (2004). On the basis of measurements made in the NW Pacific Ocean and Tokyo Bay a mean value of  $-38 \pm 4\%$  was recorded which would appear to indicate approximate equilibrium between seawater and the troposphere.

Table 1 summarises existing information on the mean  $\delta^{13}C$  value and range reported for each source.

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### 3. Kinetic isotope effects associated with known sinks for tropospheric CH<sub>3</sub>Cl

The isotopic composition of atmospheric CH<sub>3</sub>Cl can be significantly altered by the kinetic isotope effects of loss processes. Kinetic fractionations occur due to differences in reaction rates of different isotopologues in irreversible physical, chemical, or biological processes. The resulting kinetic isotope effect (KIE) associated with a loss process is defined as:

$$\text{KIE} = \frac{k_{12}}{k_{13}} \quad (2)$$

where  $k_{12}$  and  $k_{13}$  are the rate constants for loss of the lighter <sup>12</sup>CH<sub>3</sub>Cl isotopologue and the heavier <sup>13</sup>CH<sub>3</sub>Cl isotopologue, respectively. The KIE is typically expressed as a fractionation constant ( $\varepsilon$ ) in per mil (‰):

$$\varepsilon = (\text{KIE} - 1) \times 1000 \text{ ‰} \quad (3)$$

where a positive  $\varepsilon$  indicates that loss causes the remaining sample to be enriched in the heavier isotope.

The dominant loss process for atmospheric CH<sub>3</sub>Cl is via reaction with the photochemically produced OH radical in the atmosphere (Montzka and Fraser, 2003). Previously it has been assumed (Harper et al., 2001, 2003; Thompson et al., 2002) that the KIE for this reaction would be small and similar to that observed for the reaction of methane (CH<sub>4</sub>) with OH radical. The carbon KIE for reaction of CH<sub>4</sub> had been measured in the laboratory by Cantrell et al. (1990) ( $k_{12}/k_{13}=1.0054$ ,  $\varepsilon=5.4$ ) and more recently by Saueressig et al. (2001) ( $k_{12}/k_{13}=1.0039$ ,  $\varepsilon=3.9$ ). However the first measurement of the KIE for reaction of CH<sub>3</sub>Cl with OH has recently been reported by Gola et al. (2005) and reveals an unexpectedly large fractionation factor of  $\varepsilon=59\text{‰}\pm 8\text{‰}$ . The reaction of CH<sub>3</sub>Cl with chlorine radicals in the marine boundary layer constitutes another significant loss process. The KIE for this reaction is reported as  $70\text{‰}\pm 10\text{‰}$  (Gola et al., 2005).

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As explained by Thompson et al. (2002), loss of tropospheric CH<sub>3</sub>Cl to the stratosphere is caused by turbulent mixing and the transport process itself is not considered to result in an isotope fractionation. Isotope fractionation during exchange of CH<sub>3</sub>Cl between the oceans and the atmosphere has not yet been quantified, so the treatment of marine fluxes has to be approximate.

As discussed in the introduction, the microbial degradation of CH<sub>3</sub>Cl in the soil is potentially a large sink for atmospheric CH<sub>3</sub>Cl, possibly second in importance to the reaction with OH radical. Recent work by Miller et al. (2001, 2004) has shown a KIE of 1.045 and 1.049, respectively, for degradation of CH<sub>3</sub>Cl by soil bacteria. Thus a microbial sink accounting for around 20–30% of atmospheric CH<sub>3</sub>Cl significantly affects the δ<sup>13</sup>C value of atmospheric CH<sub>3</sub>Cl.

The known CH<sub>3</sub>Cl sinks in the environment and their associated fractionation factors as well as the uncertainties are presented in Table 2. Figure 1 summarises diagrammatically the atmospheric CH<sub>3</sub>Cl budget showing the carbon isotope signatures of sources and fractionation constants associated with sinks.

### 4. Budget modelling of atmospheric CH<sub>3</sub>Cl using stable carbon isotope signatures of sources and sinks

The isotopic composition of atmospheric CH<sub>3</sub>Cl reflects the weighted average isotopic signature of all the sources, and the weighted average kinetic isotope effect of all the loss mechanisms:

$$\delta^{13}\text{C}^{atm} = \sum_{i=1}^n \Phi_i^{source} \times \delta^{13}\text{C}_i^{source} + \sum_{j=1}^n \Phi_j^{sink} \times \varepsilon_j^{sink} \quad (4)$$

where δ<sup>13</sup>C<sup>atm</sup> and δ<sup>13</sup>C<sub>*i*</sub><sup>source</sup> are the carbon isotope composition of CH<sub>3</sub>Cl in the atmosphere and of the different sources *i* in per mil. Φ<sub>*i*</sub> and Φ<sub>*j*</sub> are the CH<sub>3</sub>Cl flux fraction for each source and sink. ε<sub>*j*</sub> is the fractionation constant of each sink *j* in per mil.

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To date measurements of  $\delta^{13}\text{C}$  for atmospheric  $\text{CH}_3\text{Cl}$  including marine, urban, rural, and remote sites have been reported by Rudolph et al. (1997), Tsunogai et al. (1999), and Thompson et al. (2002) with values ranging between  $-30$  and  $-47\text{‰}$ . Thompson et al. (2002) reported  $\delta^{13}\text{C}$  measurements for  $\text{CH}_3\text{Cl}$  from both Northern and Southern hemisphere. Although the number of data points from the Southern hemisphere was rather limited Thompson et al. (2002) reported a global average value  $\delta^{13}\text{C}$  of  $\text{CH}_3\text{Cl}$  of  $-36.2 \pm 0.3\text{‰}$  and this value is employed in the calculations below.

Several attempts at modelling the atmospheric  $\text{CH}_3\text{Cl}$  budget using carbon isotope ratios have been made (Harper et al., 2001, 2003; Thompson et al., 2002; Komatsu et al., 2004) but all have been hampered by the lack of data concerning isotopic fractionation of some important sources and the KIE associated with the major sinks. The recent discoveries made by Gola et al. (2005), Hamilton et al. (2003) and Keppler et al. (2004) are critically important in the development of a much refined isotopic mass balance for atmospheric  $\text{CH}_3\text{Cl}$ .

As mentioned above, the largest uncertainty in sinks is that due to microbial degradation of atmospheric  $\text{CH}_3\text{Cl}$ . In this paper we present three scenarios based on different strengths for the microbial sink and different emissions for the missing source. Data for each scenario are listed in Table 3. The values for the missing sources are calculated by subtracting total known sources from total sinks. Scenario A uses the Montzka and Fraser (2003) values for the microbial sink ( $180 \text{ Gg yr}^{-1}$ ) and the total sink ( $4005 \text{ Gg yr}^{-1}$ ) of atmospheric  $\text{CH}_3\text{Cl}$ . In scenario B a microbial sink of  $890 \text{ Gg yr}^{-1}$  (mean of estimates from Montzka and Fraser, 2003, and Harper and Hamilton, 2003) and a total sink of  $4715 \text{ Gg yr}^{-1}$  are used for the calculations. Finally, for scenario C we consider  $1600 \text{ Gg yr}^{-1}$  for the microbial sink and a total sink strength of  $5425 \text{ Gg yr}^{-1}$ .

First we determine the global average  $\text{CH}_3\text{Cl}$  composition of the sources from the atmospheric average ( $\delta^{13}\text{C}^{\text{atm}} = -36.2 \pm 0.3\text{‰}$ ) and the weighted mean for all sinks based on scenario A, B and C ( $\sum_{j=1}^n \Phi_j^{\text{sin } k} \times \varepsilon_j^{\text{sin } k} = 55.4 \pm 6.4\text{‰}$ ,  $54.2 \pm 5.5\text{‰}$  and  $53.2 \pm 4.8\text{‰}$ ).

The resulting average  $\delta^{13}\text{C}$  composition for all sources ( $\delta^{13}\text{C}_{\text{total}}$ ) is calculated to be

–91.6±6.4‰, –90.4±5.5‰ and –89.4±4.8‰ for Scenario A, B and C, respectively. Thus in all scenarios, the weighted mean of  $\delta^{13}\text{C}$  values of total sources is in the range of –92 to –89‰.

In the second step we determine the isotope composition of the missing source ( $\delta^{13}\text{C}_{\text{mis}}$ ). Using the weighted average of emissions from all known sources for which best estimates are available including biomass burning, fungi, salt marshes, tropical plants, coal combustion and incineration ( $\delta^{13}\text{C}_{\text{known sources}} = -52.5 \pm 3.6\text{‰}$ ) we can calculate the  $\delta^{13}\text{C}_{\text{mis}}$  for each scenario as follows:

$$\delta^{13}\text{C}_{\text{mis}} = \frac{[\delta^{13}\text{C}_{\text{total}} - \delta^{13}\text{C}_{\text{known sources}} \times (1 - \Phi_{\text{mis}})]}{\Phi_{\text{mis}}} \quad (5)$$

For scenario A values of  $\Phi_{\text{mis}} = 0.262$ ,  $\delta^{13}\text{C}_{\text{total}} = -91.6 \pm 6.4\text{‰}$  and  $\delta^{13}\text{C}_{\text{known sources}} = -52.5 \pm 3.6\text{‰}$  are used for calculations. Thus the isotope signature estimated for  $\delta^{13}\text{C}_{\text{mis}}$  in scenario A is  $-201.8 \pm 26.5\text{‰}$ . Using a similar approach for scenario B ( $\Phi_{\text{mis}} = 0.373$ ,  $\delta^{13}\text{C}_{\text{total}} = -90.4 \pm 5.5\text{‰}$  and  $\delta^{13}\text{C}_{\text{known sources}} = -52.5 \pm 3.6\text{‰}$ ) the estimated value for  $\delta^{13}\text{C}_{\text{mis}}$  would be  $-154 \pm 15.9\text{‰}$ . In scenario C ( $\Phi_{\text{mis}} = 0.455$ ,  $\delta^{13}\text{C}_{\text{total}} = -89.4 \pm 4.8\text{‰}$  and  $\delta^{13}\text{C}_{\text{known sources}} = -52.5 \pm 3.6\text{‰}$ ) the respective value for  $\delta^{13}\text{C}_{\text{mis}}$  is  $-133.6 \pm 11.4\text{‰}$ .

It is evident from those budget calculations that the missing global  $\text{CH}_3\text{Cl}$  source must have an highly depleted stable carbon isotope signature. Scenario A requires the missing source to have an isotopic composition which is much more depleted than any source reported to date. Based on our current knowledge it is most improbable that there is an as yet unidentified global  $\text{CH}_3\text{Cl}$  source of  $1049 \text{ Gg yr}^{-1}$  with a  $\delta^{13}\text{C}$  signature of  $-201.8 \pm 26.5\text{‰}$ . However, in scenarios B and C sources of  $1759$  or  $2469 \text{ Gg yr}^{-1}$  with respective signatures of  $-154 \pm 15.9\text{‰}$  and  $-133.6 \pm 11.4\text{‰}$  fit well with both the range of  $\text{CH}_3\text{Cl}$  emissions by senescent plants and leaf litter in the tropics and sub-tropics (Hamilton et al., 2003) and the  $\delta^{13}\text{C}$  signature of these emissions ( $\delta^{13}\text{C} = -135 \pm 12\text{‰}$ ; Keppler et al., 2004). Furthermore the source strengths of  $\text{CH}_3\text{Cl}$

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used in for scenario C are in general agreement with a three-dimensional global model study of atmospheric CH<sub>3</sub>Cl published recently by Yoshida et al. (2004). Based on model studies they have hypothesized that a missing terrestrial source of 2900 Gg yr<sup>-1</sup> is located at 30° N–30° S. Our calculations also lend credence to the proposal that CH<sub>3</sub>Cl degradation by soil microorganisms is a much bigger sink for atmospheric CH<sub>3</sub>Cl (>1 Tg yr<sup>-1</sup>) than that postulated by Montzka and Fraser (2003). While it was acknowledged above that the gross source and sink fluxes for the ocean are underestimated in these isotopic budgets because only the net exchange is considered, our calculations indicate that inclusion of substantially larger marine source and sink in the budget does not significantly influence the conclusions of this work. This remains true even if liquid/gas phase fractionations in the oceanic sources and sinks of up to 5‰ are envisaged. Only extreme fractionations would materially affect our conclusions and there is no evidence that such large isotope effects are likely.

The isotope mass balance presented here represents an important step in reducing the large uncertainties that have hitherto surrounded the atmospheric budget of CH<sub>3</sub>Cl. It provides strong support for the contention that abiotic methylation of chloride in plants and soil organic matter provides the bulk of CH<sub>3</sub>Cl released to the atmosphere. Further refinement of this budget will require accurate measurement of source strengths in the field and their respective isotopic signatures.

**Acknowledgements.** We thank the European Commission for a Marie Curie-Research Training Grant (MCFI-2002-00022) awarded to F. Keppler and the research group of C. Nielsen for providing the exciting data of the KIE measurements. The ISOSTRAT project in Heidelberg was funded by the BMBF within the AFO2000 project (grant 07ATC01).

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**Table 1.** Known sources of tropospheric CH<sub>3</sub>Cl and corresponding carbon isotope signatures.

Sources	Source (best estimate) <sup>a</sup> (Gg yr <sup>-1</sup> )	Source (full range) <sup>a</sup> (Gg yr <sup>-1</sup> )	Carbon isotope signature (‰)	Range in δ <sup>13</sup> C (‰)
<i>Terrestrial</i>				
Biomass burning	911 <sup>b</sup>	655 to 1125	-47 <sup>c</sup>	12
Tropical plants	910	820 to 8200	-71 <sup>d</sup>	2
Fungi	160	43 to 470	-43 <sup>e</sup>	2
Salt marshes	170	65 to 440	-62 <sup>d,f</sup>	3
Wetlands	40	6 to 270	?	
Coal combustion	105	5 to 205	-47 <sup>g</sup>	12
Incineration	45	15 to 75	-47 <sup>g</sup>	12
Industrial	10		-51 <sup>h</sup>	
Rice	5		?	
Decay of organic matter in topsoil	?	?	?	
Senescent and leaf litter	?	30 to 2500	-135 <sup>i</sup>	12
<i>Marine</i>				
Oceans	600	325 to 1300	-38 <sup>j</sup>	4
Total sources	2956 + ?	1964 to 14 585		

<sup>a</sup> Values for source (best estimate) and source (full range) were taken from Montzka and Fraser (2003), except for emissions associated with senescent and leaf litter which are from Hamilton et al. (2003).

<sup>b</sup> Although lower estimates for CH<sub>3</sub>Cl emissions from biomass burning have been reported by Andreae and Merlet (2001) and Yoshida et al. (2004) the value in the Table is used in our model.

<sup>c</sup> Thompson et al. (2002)

<sup>d</sup> Harper et al. (2003) and unpublished results (Hamilton, Yokouchi, Yukawa and Harper)

<sup>e</sup> Harper et al. (2001)

<sup>f</sup> Bill et al. (2002)

<sup>g</sup> based on the assumptions made in the text

<sup>h</sup> mean of values reported by Rudolph et al. (1997), Holt et al. (1997) and Harper et al. (2001)

<sup>i</sup> Keppler et al. (2004)

<sup>j</sup> Komatsu et al. (2004)

? denotes that no value has been provided

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**Table 2.** Known sinks of tropospheric CH<sub>3</sub>Cl and the mean fractionation factors reported for each.

Sinks	Sink (best estimate) <sup>a</sup> (Gg yr <sup>-1</sup> )	Sink (full range) <sup>a</sup> (Gg yr <sup>-1</sup> )	Kinetic isotope effect (KIE) ε (‰)	Error in ε (‰)
Reaction with OH in troposphere	-3180	-2380 to -3970	59 <sup>b</sup>	8
Loss to stratosphere	-200	-100 to -300	0 <sup>c</sup>	5
Reaction with Cl in marine boundary layer	-370	-180 to -550	70 <sup>b</sup>	10
Microbial degradation in soil	-180 (-1600)	-100 to -1600	47 <sup>d</sup>	3
Loss to polar cold ocean waters	-75	-37 to -113	?	?
Total sinks	-4005 (-5425)	-2797 to -6533		

<sup>a</sup> Values for sink strength (best estimate and full range) were taken from Montzka and Fraser (2003), except for the value shown in brackets for the sink (best estimate) for microbial degradation in soil which is from that suggested by Harper and Hamilton (2003) and discussion in the text of this manuscript.

<sup>b</sup> Gola et al. (2005)

<sup>c</sup> Thompson et al. (2002) and discussion in this manuscript

<sup>d</sup> mean of Miller et al. (2001) and Miller et al. (2004)

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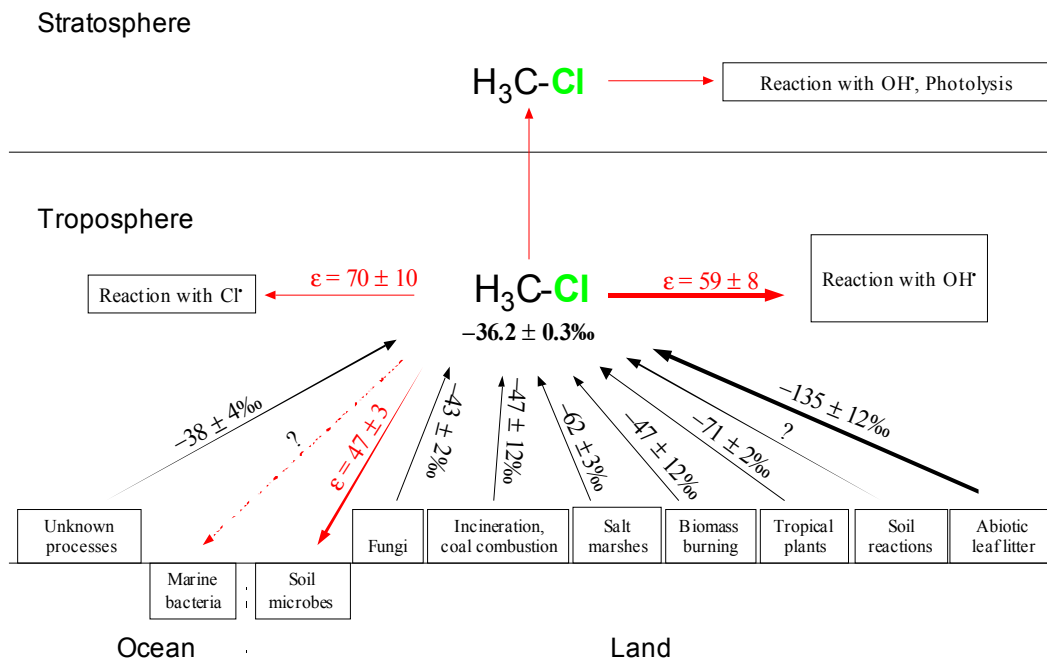
**Table 3.** Source and sink strengths of CH<sub>3</sub>Cl<sup>a</sup> used for three scenarios.

Scenario	1	2	3
<i>Total sources</i>	4005	4715	5425
Biomass burning	911	911	911
Tropical plants	910	910	910
Fungi	160	160	160
Salt marshes	170	170	170
Wetlands	40	40	40
Coal combustion	105	105	105
Incineration	45	45	45
Industrial	10	10	10
Rice	5	5	5
Oceans	600	600	600
Total known sources	2956	2956	2956
Missing source	1049	1759	2469
<i>Total sinks</i>	−4005	−4715	−5425
Reaction with OH in troposphere	−3180	−3180	−3180
Loss to stratosphere	−200	−200	−200
Reaction with Cl in marine boundary layer	−370	−370	−370
Loss to cold ocean waters	−75	−75	−75
Microbial degradation in soil	−180	−890	−1600

<sup>a</sup> Units are in Gg yr<sup>−1</sup>

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**Fig. 1.** Scheme of major sources and sinks involved in the global  $\text{CH}_3\text{Cl}$  cycle and the corresponding carbon isotope signatures and fractionation factors. Black arrows show sources and red arrows indicate sinks of  $\text{CH}_3\text{Cl}$  in the environment.

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